Reproductive consequences of developmental phytoestrogen exposure

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Abstract

Phytoestrogens, estrogenic compounds derived from plants, are ubiquitous in human and animal diets. These chemicals are generally much less potent than estradiol but act via similar mechanisms. The most common source of phytoestrogen exposure to humans is soybean-derived foods that are rich in the isoflavones genistein and daidzein. These isoflavones are also found at relatively high levels in soy-based infant formulas. Phytoestrogens have been promoted as healthy alternatives to synthetic estrogens and are found in many dietary supplements. The aim of this review is to examine the evidence that phytoestrogen exposure, particularly in the developmentally sensitive periods of life, has consequences for future reproductive health.

Introduction

 Phytoestrogens are plant compounds that are structurally similar to estradiol and can interact with estrogen receptors (ERs) to promote and/or inhibit estrogenic responses. These compounds are ingested as part of a normal diet and are frequently included in plant-based dietary supplements such as those advertised to alleviate symptoms of menopause. The major phytoestrogen groups are isoflavones (genistein, daidzein, glycitein, and formononetin), flavones (luteolin), coumestins (coumestrol), stilbenes (resveratrol), and lignans (secoisolariciresinol, matairesinol, pinoresinol, and lariciresinol) (Moutsatsou 2007). Isoflavones are found at high concentrations in soybean products whereas lignans are found in flax seed, coumestans are found in clover, and stilbenes are found in cocoa- and grape-containing products, particularly red wine. The β-D-glycoside form of genistein – genistin – makes up the majority (55–65%) of the isoflavone content in soy products (Setchell et al. 1997). The β-D-glycoside form of daidzein – daidzin – comprises about 30–35%, and glycitin, glycitein, biochanin A, and formononetin together account for <10% of soy isoflavones. For additional reference regarding dietary isoflavone content, see the USDA-Iowa State University Database on the Isoflavone Content of Foods (release 1.3, U.S. Department of Agriculture, Beltsville, MD, USA) at http://www.nal.usda.gov/fnic/foodcomp/Data/isoflav/isoflav.html and Thompson et al. (2006).

Several decades ago, observational studies raised concerns regarding the reproductive toxicity of phytoestrogens consumed in the diet. The earliest evidence that naturally occurring phytoestrogens could cause reproductive disturbances in mammals was reported in 1946, indicating that sheep that grazed on red clover were infertile due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984). Finally, a population of captive cheetahs exhibited infertility due to the estrogenic content in the clover (Bennetts et al. 1946, Morley et al. 1966). About 20 years later, a similar observation was made in cows that had fertility disturbances resulting from periods of stall-feeding on red clover (Kallela et al. 1984).
signaling (Young et al. 1964, MacLusky & Naftolin 1981, Park & Jameson 2005, Ma 2009, Sakuma 2009). For this reason, endocrine-disrupting chemicals, including phytoestrogens, could have a significant impact on development in ways that affect later reproductive health. Furthermore, these effects could have long-term consequences for the offspring of the affected individuals.

Since the initial observations of adverse effects of phytoestrogens on adult mammalian reproductive health, numerous studies have been performed in laboratory animal models to evaluate the estrogenic activity of phytoestrogens (Diel et al. 2000, Jefferson et al. 2002) and their effects on development and reproduction. Most studies have been performed in rodent models, with various phytoestrogens administered either prenatally to the pregnant dam or postnatally to the pups (Burroughs et al. 1995b, Nagao et al. 2001, Newbold et al. 2001, Kouki et al. 2003, Nikaido et al. 2004, Jefferson et al. 2005, Bateman & Patisaul 2008). Together, these studies provide a wealth of evidence that exposure of developing animals to phytoestrogens negatively impacts future fertility. We present here background information regarding human phytoestrogen exposure that provides the rationale for studies in animal models and briefly review the mechanisms of phytoestrogen action and the critical concept of specific windows of developmental sensitivity to disruption. We then review in detail the information regarding phytoestrogens and reproductive health outcomes, focusing mainly on studies of developmental phytoestrogen exposure in animal model systems, and concluding with available information from human studies.

Human phytoestrogen exposure

The amount of phytoestrogens consumed in the diet is highly variable and depends primarily on soy consumption; the resulting serum circulating levels of phytoestrogens are also highly variable (Verkasalo et al. 2001). For example, the highest circulating levels of genistein are seen in Asian populations who traditionally consume high levels of soy; Japanese men and women have ∼500 nM genistein in circulation (Adlercreutz et al. 1994, Morton et al. 2002). There is also a large range of soy consumption across Europe. One study found mean serum genistein levels of 148 nM in a region with many vegetarians and much lower mean levels (2.6–22.6 nM) in other regions where fewer vegetarians lived (Peeters et al. 2007). This variability was also seen in another European study in which subjects were divided into four groups based on the amount of reported dietary soy intake (lowest to highest) and the corresponding serum genistein levels were 14.3, 16.5, 119, and 378 nM (Verkasalo et al. 2001). Serum genistein levels in Finnish omnivores and vegetarians were 4.9 and 17.1 nM, respectively (Adlercreutz et al. 1994). Americans, in general, have low levels of genistein because soy is not prevalent in the diet. A nonrepresentative subset of samples from the National Health and Nutrition Examination Survey showed an average serum genistein level of 17.4 nM, with about half of the samples tested having nondetectable levels (Valentin-Blasini et al. 2003).

Because of maternal dietary phytoestrogen consumption, human fetuses are exposed to phytoestrogens during in utero development (Franke et al. 1998, Foster et al. 2002). The fetus is exposed to substances in maternal serum through the filter of the placenta. Although in theory the placenta serves as a barrier to protect the fetus from exposure to harmful chemicals, in fact most drugs and environmental chemicals enter the fetal circulation by either passive diffusion or active transport (Syme et al. 2004). An important modifier of fetal exposure is binding of chemicals to maternal serum proteins. For example, human steroid hormone-binding protein (SHBG) binds estradiol with high affinity and likely protects the fetus from the effects of maternal estradiol on the brain and reproductive tract. However, human pregnancy plasma, which includes SHBG, has negligible affinity for phytoestrogens (Milligan et al. 1998); so fetal exposure to phytoestrogens is directly related to the maternal serum circulating level. The fetal counterpart to serum albumin is α-fetoprotein (AFP), the major serum constituent in the fetus. In rodents, AFP efficiently binds estradiol and could also serve to protect the fetus from its developmental effects. However, human AFP does not bind estradiol (Keel et al. 1992), and neither type of AFP binds phytoestrogens, and so the presence of AFP in human fetal serum is not protective. Indeed, genistein and daidzein are detected in amniotic fluid from the second trimester of pregnancy at levels similar to those observed in adult serum, and 10- to 20-fold higher than the average amniotic fluid estradiol levels at that time of pregnancy (Robinson et al. 1977, Foster et al. 2002).

Although human infants can be exposed to phytoestrogens via breast milk (Franke et al. 1998), the highest human exposure to phytoestrogens occurs in infants consuming soy-based infant formulas (Setchell et al. 1997, Cao et al. 2009). Infants fed exclusively soy-based formula have serum genistein levels of 1–10 μM when tested at random time intervals after feeding (Cao et al. 2009). Differences in the time of serum collection after feeding may explain the ∼10-fold difference in values from different individuals. The estrogenic activity of soy-based formula in human infants was documented recently in a largely cross-sectional study of the plausible effects of estrogen exposure on male and female infants fed breast milk, soy formula, or cow milk formula (Bembaum et al. 2008). Vaginal/introital wall cells in all the girls showed estrogen effects at birth, which were largely lost by about 3 months. The two girls fed soy formula and studied at 6 months showed re-estrogenization (Fig. 1). This small study is consistent with an estrogen...
Relevance of animal models to human exposure levels

Ideally, studies to examine the effects of phytoestrogens on humans could be conducted in human subjects. However, there is wide variation in human exposures, these exposures are difficult to measure accurately, and the exposures are inherently difficult to control effectively (Verkasalo et al. 2001). There is also extensive variability in the phytoestrogen content of many dietary sources over time, whether standard food products or commercial botanical extracts that are sold as dietary supplements (Thompson et al. 2006). Furthermore, not all humans metabolize phytoestrogens in the same way because of differences in the activity of metabolizing enzymes and the influence of gut microflora on phytoestrogen bioavailability (de Cremoux et al. 2010). Together, these factors complicate both design and interpretation of human studies, and combined with the ethical issues regarding experimentation in humans, have led to extensive reliance on studies that utilize animal models.

The relevance to human health of studies performed in animal models has been questioned because in many of the animal studies exposure to phytoestrogens was by a non-oral route, whereas most human phytoestrogen exposure is due to dietary intake (Verkasalo et al. 2001). Non-oral exposures are frequently chosen for rodent models of developmental phytoestrogen exposure because the size difference between rodents and humans, particularly in the neonatal period, precludes achieving similar serum phytoestrogen levels by an oral exposure route. Given that the serum concentration determines the amount of bioactive chemical to reach a target tissue, we believe that the serum measurement is the most important parameter to consider when attempting to model human exposures. This idea has been substantiated for genistein in two recent publications where direct comparisons of biological effects were made after exposure via different routes and forms of the chemical (Jefferson et al. 2009a, Cimafranca et al. 2010).

Pharmacokinetic analyses of s.c. exposure of neonatal mice to genistein and oral exposure to genistin or genistein demonstrated that oral exposure to genistin results in peak circulating levels of total genistein that are higher than those following s.c. injection, but the total area under the curve over a 24-h period is very similar (Fig. 2; Jefferson et al. 2009a). Both exposure methods result in similar biological effects, suggesting that both peak and sustained genistein levels contribute to biological outcomes. In contrast, following an orally administered dose of genistein, there is ~90% less genistein found in circulation when compared with the same dose of genistein administered s.c., and about a 20-fold lower peak genistein level (Jefferson et al. 2009a). Genistein also has much lower bioavailability than genistin when given orally to adult rats (Kwon et al. 2007). These data confirm that oral genistein dosing results in significantly lower tissue bioavailability when compared with both oral genistin exposure and s.c. genistein exposure. Furthermore, both oral genistin and s.c. genistein dosing in this neonatal mouse model system result in low micromolar serum circulating total genistein levels that closely approximate the levels measured in infants fed soy-based infant formula (Fig. 2; Cao et al. 2009, Jefferson et al. 2009a). These findings support their use in modeling developmental exposure of human infants to phytoestrogens.

Mechanisms of phytoestrogen action

The mechanisms of action of various phytoestrogens have been reviewed recently (Dang 2009, Li & Tollefsbol 2010, Lorand et al. 2010, Shanle & Xu 2011) and therefore will only be mentioned briefly here. As the name implies, phytoestrogens can interact with the classical estrogen receptors, ERα and ERβ, which mediate many of their downstream actions. Their affinities for ERα and ERβ are relatively weak compared with estradiol, and they can have agonist or antagonist activity depending on whether estradiol is also present (Shanle & Xu 2011). This is in contrast to the well-known endocrine-disrupting chemical diethylstilbestrol (DES),
which has higher affinity than estradiol for the ERs that mediate the majority of its actions (Korach & McLachlan 1985, Shanle & Xu 2011). Several phytoestrogens are selective ER modulators that have greater affinity for ERβ than ERα (Lorand et al. 2010). However, phytoestrogens and xenoestrogens such as DES also affect numerous other signaling pathways including nongenomic signaling mediated by oxidative stress pathways, tyrosine kinases, nuclear factor-κB, and extracellular signal-regulated kinases (Watson et al. 2007, de Souza et al. 2010). In addition to classical ERs, phytoestrogens serve as ligands for peroxisome proliferator-activated receptors, the nonclassical ER GPER1 (previously GPR30), the estrogen-related receptors, and the aryl hydrocarbon receptor (Suetsugi et al. 2003, Dang 2009, Prossnitz & Barton 2009, de Souza et al. 2010). Besides direct actions to modulate signaling pathways, phytoestrogens can alter epigenetic marks by altering the activities of DNA and histone methyltransferases, NAD-dependent histone deacetylases, and other modifiers of chromatin structure (Labinskyy et al. 2006, Li & Tollefsbol 2010, Shanle & Xu 2011). Finally, phytoestrogens can competitively inhibit the production of estradiol by aromatase (Kao et al. 1998, Shanle & Xu 2011), which would lead to lower endogenous estrogen levels. The complexity of phytoestrogen actions is increased by the fact that these compounds are frequently present in vivo as mixtures of several dietary components that can affect multiple signaling pathways or affect the same pathways in opposing directions.

Windows of developmental sensitivity

The fetus and neonate are highly sensitive to environmental chemical exposures because organogenesis, rapid growth, and extensive tissue differentiation occur during these developmental periods and therefore small perturbations can have important consequences. In addition, metabolic processing and elimination mechanisms are immature in the fetus and neonate, and so detoxification is inefficient (Beath 2003, Chen et al. 2006). Furthermore, many developmental processes are dependent on steroid hormones and secreted proteins whose activities have the potential to be altered by phytoestrogens based on the mechanisms of action outlined earlier. For example, testosterone secreted by the fetal testis is a key mediator of male gonad and reproductive tract development (Park & Jameson 2005). Fetal exposure to estrogenic compounds can suppress testosterone synthesis, leading to cryptorchidism and adult testis dysfunction (Clark & Cochrum 2007, Wohlfahrt-Veje et al. 2009). In addition, exposure to estrogenic chemicals during fetal life disrupts female reproductive tract development by altering the expression of genes encoding secreted signaling proteins critical for directing this process (Ma 2009); these effects have permanent consequences for reproductive tract morphology and function in both rodents and humans (Newbold et al. 2004, Baird & Newbold 2005).

Although most organ systems in the fetus are highly sensitive to developmental insults, this sensitivity persists during postnatal life in several specific organ systems because they continue to develop after birth. For example, in humans, the brain continues to undergo significant growth and differentiation during the first several years of life, and there is evidence of continued brain development through late adolescence (reviewed in Sisk & Zehr 2005). Similarly, in animals including...
humans, the female reproductive tract is not completely differentiated at birth, but continues to undergo cellular differentiation until just before the onset of puberty (Valdes-Dapena 1973, Gray et al. 2001). Many of these postnatal differentiation events are dependent, at least in part, on steroid hormone signaling. For this reason, endocrine-disrupting chemicals, including phytoestrogens, could have a significant impact on postnatal development in ways that affect later reproductive health.

Phytoestrogen exposure and reproductive health in animal models

Reproductive health critically depends on proper sexually dimorphic fetal and postnatal development followed by sex-specific and coordinated functioning of the brain, gonads, reproductive tract, and external genitalia. Although other organ systems, particularly endocrine-related organs, also contribute to reproductive health, there is little information available regarding the impact of phytoestrogens on these other organs as they relate to reproductive function. For this reason, we limit the discussion that follows to the major organs and essential processes considered to be part of the reproductive system.

Estrous cycle

We have used a mouse model of neonatal exposure to the phytoestrogen genistein to study the impact of physiologically relevant levels of phytoestrogens on female reproductive health (Jefferson et al. 2005, 2009b). The initial dosing strategy achieved serum levels ranging from those observed in humans eating a nonvegetarian diet all the way to levels measured in infants fed soy-based infant formula as their major food source. Neonatal genistein treatment causes abnormal estrous cycles and anovulation as a result of abnormalities in the hypothalamic–pituitary–ovarian axis function (Jefferson et al. 2005). In the rat, neonatal genistein exposure alters pituitary sensitivity to a GnRH challenge, with a low genistein dose causing increased sensitivity and higher doses causing decreased sensitivity as indicated by LH release (Faber & Hughes 1993). More recent work demonstrated that neonatal genistein alters hypothalamic kisspeptin signaling pathways and GnRH activity (Fig. 3; Losa et al. 2011). These pathways regulate both timing of pubertal onset and estrous cyclicity, and thus provide a likely explanation for the abnormal estrous cyclicity and anovulation observed in the mouse model. Based on studies in adult rats, resveratrol can also disrupt estrous cyclicity, although the underlying mechanism is not clear (Henry & Witt 2002).

Sexually dimorphic behavior

Pre- and postnatally, the brain develops in a sexually dimorphic fashion that contributes to differential reproductive behaviors in response to hormonal stimuli during adulthood (Simerly 2002, Sakuma 2009, Roselli & Stormshak 2010, Henley et al. 2011). For example, the sexually dimorphic nucleus of the preoptic area (SDN) is substantially larger in males than in females; in rams and rats, the size of this region is associated with sexual partner preference (Roselli & Stormshak 2010, Henley et al. 2011). The SDN size difference is a result of exposure to testosterone, generated by the testis, and estradiol, which is generated in the brain by aromatization of testosterone. Indeed, testosterone and estradiol serve as the most important factors to establish permanent sex differences in brain organization during the fetal and neonatal periods (reviewed in Arnold (2009)). Although most sex differences are likely established during prenatal and neonatal development, it was shown recently in the rat that new cells are added to sexually dimorphic nuclei during adolescence in response to steroid hormone treatments (Ahmed et al. 2008), demonstrating the long-term sensitivity of sexually dimorphic brain regions to steroid hormone-mediated signaling.

Several phytoestrogens affect brain development and sexual behavior in animal models. For example, exposure of male rat pups to resveratrol via nursing is associated with smaller SDN size and reduced sociosexual behavior in adulthood (Henry & Witt 2006). Coumestrol administration to neonatal female rats markedly reduced lordosis behavior in adulthood (Kouki et al. 2005). Finally, neonatal genistein exposure affected sexual behavior of adult male rats, an effect likely mediated by ERβ (Wisniewski et al. 2003, Sullivan et al. 2011). Notably, all these exposures occur during the neonatal period but affect adult behaviors,
highlighting the importance of the timing of exposures in determining specific long-term outcomes.

**Testis function**

There is some evidence in animal models that developmental exposure to phytoestrogens affects testis function, but many studies found no effects, and so it is difficult to draw definitive conclusions (Cederroth et al. 2010). Several studies that did show alterations in testis function used long-term exposure protocols, beginning during gestation and continuing through adulthood. For example, two studies demonstrated that male rats chronically exposed to genistein had abnormalities in spermatogenesis (Delclos et al. 2001, Eustache et al. 2009). One of these studies also demonstrated that genistein caused alterations in sperm motility and a reduction in litter size accompanied by evidence of post-implantation embryo loss when the adult rats underwent fertility testing (Eustache et al. 2009). In a set of very nicely designed twin studies in marmosets, neonatal soy milk exposure was shown to alter testis cellular morphology and reduce serum testosterone levels when compared with the same parameters in twin siblings fed standard milk formula, but fertility was not different between the two groups (Sharpe et al. 2002, Tan et al. 2006).

**Ovarian function**

Female germ cells complete proliferation and begin meiosis prenatally and then arrest in meiosis I before or at the time of birth. Prenatal development is absolutely critical for female reproductive health because this is the last time that new daughter germ cells are generated. In rodents, primordial follicles form during the first postnatal week by the migration of pregranulosa cells to surround individual oocytes as oocyte nests break down. Exposure to genistein during this process causes failure of oocyte nest breakdown and leads to the development of multi-oocyte follicles (Chen et al. 2007, Cimafranca et al. 2010, Losa et al. 2011). It does not appear that the occurrence of multi-oocyte follicles impacts fertility; however, because embryos derived from neonatal genistein-treated mice are fully competent to develop to term if transferred into pseudopregnant control mice (Jefferson et al. 2009b).

**Female reproductive tract function**

In newborn rodents, the female reproductive tract is a simple bifurcated tube comprised of an outer epithelium, a muscle layer, a thin stromal layer, and a simple columnar epithelium. Extensive gross morphological and cellular differentiation begins in the immediate postnatal period and is complete by about postnatal day 15 in the mouse (Gray et al. 2001). This differentiation process is under the regulatory control of homeobox transcription factors, particularly the posterior Hoxa genes, and secreted signaling molecules that induce formation of the morphologically distinct reproductive tract regions: oviduct, uterus, cervix, and vagina (Ma 2009, Umezu et al. 2010). In addition to regional differentiation, the endometrial epithelium undergoes glandular morphogenesis postnatally to form branched glandular structures in the endometrial stroma. These cellular and regional differentiation processes are exquisitely sensitive to disruption by steroid hormone signaling (Gray et al. 2001).

Phytoestrogens dramatically alter development of the rodent female reproductive tract. Neonatal genistein exposure disrupts oviductal morphogenesis by altering hedgehog signaling pathways and expression of homeobox genes and other transcription factors (Jefferson et al. 2011). As a result, the oviduct becomes ‘posteriorized’, i.e. expresses genes and proteins normally observed only in posterior regions of the reproductive tract (cervix and vagina) (Fig. 4). The abnormally expressed genes include several developmentally important homeobox transcription factors that are absent in control oviducts, including Pitx1 and Six1 (Jefferson et al. 2011). In addition, microarray analysis of oviduct gene expression on pregnancy day 2 revealed that neonatal genistein exposure causes significant alterations in gene expression in the category of inflammatory response pathways, many of which are hormonally regulated. This suggests the presence of alterations in oviductal mucosal immune responses during early pregnancy that could impact embryo survival and development (Jefferson et al. 2011).

There is a wealth of information available regarding the effects of developmental exposure to phytoestrogens on uterine morphology and function. For example, treatment of immature rats for 1 week with resveratrol increased uterine wet weight and glandular morphogenesis (Singh et al. 2011). These changes were accompanied by alterations in several endometrial prosurvival or antiapoptotic factors, but whether resveratrol treatment at this time in development has permanent effects is unknown. Treatment of neonatal rats with coumestrol caused early vaginal opening and an initial increase in uterine wet weight during the time of treatment followed by a later decrease in adult uterine weight (Medlock et al. 1994, Losa et al. 2011). This was also true for mice treated neonatally with DES (Newbold et al. 2004). When coumestrol was administered to rats slightly later, during postnatal days 10–14 that are critical for glandular morphogenesis, fewer endometrial glands were observed in the adults and ER expression was reduced (Medlock et al. 1994, 1995a). Mice treated neonatally with coumestrol developed squamous metaplasia and had abnormal collagen deposition in the uterine wall by 22 months of age.
Together, these studies overwhelmingly support the concept that brief exposure to phytoestrogens during the developmentally sensitive periods can have permanent effects on uterine morphology and function.

Although we are mainly focusing here on reproductive outcomes, developmental phytoestrogen exposure is also associated with uterine carcinogenesis. Indeed, mice treated neonatally with genistein had a 35% incidence of uterine cancer by 18 months of age (Newbold et al. 2001). This incidence is similar to the incidence seen when mice are given an equally estrogenic dose of DES, suggesting that it is mediated by ER signaling. Of note, the cancer phenotype is not observed if the DES-exposed mice are ovariectomized before puberty, indicating that a ‘second hit’ of steroid hormone exposure is required (Newbold et al. 1990). In contrast, a recent study demonstrated the beneficial effects of neonatal exposure to phytoestrogens on uterine carcinogenesis (Begum et al. 2006). This study used mice heterozygous for a knockout allele of the tumor suppressor phosphatase and tensin homolog (Pten) that have a very high incidence of endometrial hyperplasia and adenocarcinoma by 12 months of age. Neonatal exposure to genistein at the same doses used by Newbold et al. (2001) strongly suppressed endometrial carcinogenesis in the Pten<sup>−/+</sup> model. Of note, the genistein-exposed Pten<sup>−/+</sup> mice demonstrated alterations in endometrial stromal morphology and reductions in expression of Hoxa10 and Hoxa11 similar to those we had observed after treating wild-type mice with genistein (W NJ and C JW, unpublished observations), suggesting
that the different outcome was related to differences in cancer etiology in the two models and not differences in genistein effects.

The mechanisms underlying the persistence of altered uterine phenotypes into adulthood after these developmental exposures, and in the case of neonatal DES-exposed mice, transmission to subsequent generations (Newbold et al. 1998, 2000), are not known and are active areas of current research. The most likely explanation is that endocrine disruption during development results in permanent epigenetic changes responsible for the abnormal phenotype that are maintained into adulthood and can be stably transmitted to the next generation. There are precedents for the epigenetic changes occurring after developmental exposure to endocrine disruptors including phytoestrogens and other environmental cues (Wade & Archer 2006, Jirtle & Skinner 2007, Jones & Baylin 2007). These exposures can alter the DNA methylation status of specific genes (Li et al. 2007, Tang et al. 2008) or alter methylation patterns on a more global scale (Anway et al. 2005, Dolinoy et al. 2007), suggesting that methylation is one potential mechanism for this phenomenon. Indeed, there is some evidence in mice that prenatal DES exposure can increase methylation of Hoxa10, a gene important for uterine development and adult uterine function (Bromer et al. 2009). However, the functional significance of these methylation changes is not known, and similar alterations in Hoxa10 methylation were not observed after prenatal genistein or daidzein exposure (Akbas et al. 2007). A recent study demonstrated that exposure of ovariec-tomized adult rats to high doses of genistein caused subtle decreases in methylation of the steroidogenic factor 1 promoter associated with increased expression of this gene in the endometrium, but whether the protein level was altered and whether the effects were permanent is unknown (Matsukura et al. 2011). We anticipate that future studies that could examine a more complete scope of epigenetic alterations such as histone modifications and alterations in three-dimensional chromatin structure will help to shed light on the mechanistic basis of the permanent reproductive tract alterations observed after developmental phytoestrogen exposure.

Reproductive tract influence on embryo development in the mouse

It should come as no surprise that abnormalities in reproductive tract development can result in alterations in embryo development. Adult females treated neonatally with genistein occasionally ovulate spontaneously up to about 8 weeks of age, but after that time remain in a state of persistent estrous and are anovulatory (Jefferson et al. 2005). They can, however, be superovulated with gonadotropins and then bred to fertile males. These mice achieve pregnancy, but fertilization is delayed by several hours when compared with controls (Jefferson et al. 2009b).

If the pronuclear-stage embryos are removed from the oviduct and cultured in vitro, they develop normally to the blastocyst stage and can develop to term if transferred into control pseudopregnant females. If left in the oviduct, however, by the third day of pregnancy, about half of the embryos are reabsorbed (Fig. 5). The surviving embryos develop in vivo at a faster rate than in the case of controls and have an altered trophectoderm: inner cell mass ratio (W N J and C J W, unpublished observations). Some of the embryos that survive transit through the oviduct implant in the uterus, but the implantation sites are abnormal and the embryos do not develop to term (Jefferson et al. 2005). This is not a consequence of an abnormal steroid hormone milieu because the serum levels of estrogen, progesterone, and testosterone in these mice are within the normal range on days 6, 8, and 10 of pregnancy (Jefferson et al. 2005). Instead, the uterus is not competent to support embryo development, even if blastocysts from untreated control mice are transferred into pseudopregnant genistein-treated mice (Jefferson et al. 2009b). These findings demonstrate that neonatal phytoestrogen exposure alters female reproductive tract development in ways that can directly affect the development of the subsequent generation of embryos.

**Figure 5** Embryo loss during transit through the oviduct in adult mice treated neonatally with genistein. Embryos were flushed from the oviduct and uterus of control (white bars) and neonatal genistein-treated (black bars) superovulated mice at the indicated times after human chorionic gonadotropin (hCG) administration and mating. Graph shows mean number±S.E.M. of embryos per mouse. *P<0.05 compared to control at the same time point. Schematic above the graph indicates expected embryo stage and location within the female reproductive tract at the different time points. Modified from Jefferson et al. (2009b).
External genitalia

Male differentiation of the external genitalia occurs prenatally in response to testosterone produced by the fetal testis and dihydrotestosterone generated locally in external genital tissues (Park & Jameson 2005). Several studies have demonstrated the effects of phytoestrogens on external genital development or function in rodents. Gestational and/or lactational genistein exposure can cause a reduction in anogenital distance in male mice and rats (Levy et al. 1995, Wisniewski et al. 2003, 2005, Ball et al. 2010). One study demonstrated that erectile function is abnormal in juvenile rats exposed to daidzein for 3 months (Pan et al. 2008). Finally, hypospadias was observed in 25% of male mice exposed prenatally to genistein (Vilela et al. 2007). This finding is likely explained by genistein-mediated alterations in the expression of genes important for urethral morphogenesis including some regulated by ER signaling (Ross et al. 2011).

In the absence of prenatal androgen exposure, female external genitalia will form. Completion of female mouse external genitalia development is delayed in comparison with males because at birth there is incomplete fusion of the urethral folds. Urethral fold fusion is not complete until after postnatal day 5 and is sensitive to disruption by estrogenic chemicals in the neonatal period (Miyagawa et al. 2002). We recently demonstrated that female mice exposed neonatally to genistein have a 100% incidence of hypospadias (Padilla-Banks et al. 2011). As in males, exposure of female rats during gestation to phytoestrogens including genistein and lignans can cause reduced anogenital distance (Tou et al. 1998, Delclos et al. 2009).

Phytoestrogen exposure and reproductive health in humans

Although work in animal model systems suggests that phytoestrogens impact reproductive health, we next consider evidence in humans that this might be the case. Human studies on this topic are difficult to carry out given their typical complexity and dependence on relatively long-term observational studies rather than the treatment-based outcomes measured in animal models. Nevertheless, several studies have been published regarding the effects of developmental phytoestrogen exposure on the brain, reproductive tract, and external genitalia that are relevant to human reproduction.

Sexually dimorphic behavior

Studies in animal models have indicated that many of the effects of androgens on behavior are mediated by activation of ER-based signaling, presumably by aromatization of androgens to generate estrogen within the brain (Wu & Shah 2011). It is well established in humans that prenatal exposure to androgens can result in masculinization of behavior (Hines 2008). There is much less information available regarding the effects of estrogens or phytoestrogens on sexually dimorphic behavior in humans. However, a longitudinal study examining the impact of human infant exposure to soy-based formula on gender-role-play behavior was recently published (Adgent et al. 2011b). This study of over 7000 children enrolled in the Avon Longitudinal Study of Parents and Children (ALSPAC) found an association between early-life soy exposure (starting before age 4 months) and masculinized play behavior in girls at age 42 months. There were no measurable effects in boys, and the effects on girls were attenuated by 57 months of age. The consistency of these findings with behavioral changes observed in well-controlled studies using animal models suggest that evidence obtained from animal studies regarding the effects of exposure to estrogenic chemicals in early life on adult reproductive behavior is relevant to behavioral outcomes in humans.

Puberty

The timing of pubertal onset is a result of complex interactions between endocrine system organs including the brain, gonads, and adipose tissue that depend on peptide and steroid hormone actions, and as such could be affected by early-life phytoestrogen exposure. A subset of girls in the ALSPAC cohort described earlier who provided information regarding pubertal timing (n=2920) was used to examine associations between early-life soy exposure and timing of menarche (Adgent et al. 2011a). The early-soy-exposure group had a 25% increased risk of menarche before age 12 when compared with girls fed other types of infant formula or milk. However, the study was limited by the fact that only 2% of the subjects had early soy exposure, and loss to follow-up was significant.

A recent case–control study of ~200 Korean girls aged 8–10 years demonstrated that high serum isoflavone levels (daidzein and genistein) are associated with central precocious puberty (Kim et al. 2011). Central precocious puberty was diagnosed using strict criteria of breast budding before age 8, bone age advanced more than 1 year, and a positive GnRH stimulation test. In fact, there was about a fourfold increased risk of central precocious puberty in girls who had total isoflavone levels in the middle (30–70 nM) or high (≥70 nM) range compared with the lowest range (<30 nM). This study is in contrast to two observational studies of the timing of pubertal onset that demonstrated an association of higher urinary isoflavone levels with slightly later onset of breast development (Wolff et al. 2008, Cheng et al. 2010). One of these studies was in an ethnically mixed population of 9-year-old girls in New York City and the other used a prospectively followed cohort of children in
Germany. A multicenter prospective longitudinal study of over 1100 girls in the USA found essentially no association between urinary phytoestrogen levels and onset of breast development (Wolff et al. 2008). Although there were many methodological differences between the studies, one interpretation of the slightly conflicting results is that phytoestrogen exposure early in the prepubertal period when endogenous estradiol levels are low could predispose girls to central precocious puberty, whereas later phytoestrogen exposure could serve to interfere with endogenous estrogen-mediated stimulation of breast development because of weak ER agonist activity.

**Female reproductive tract**

The consequences of exposing pregnant women to DES made clear the potential impact of estrogenic chemicals on female reproductive tract development in humans. In addition to the increased incidence of vaginal cancer in daughters of DES-treated women, there is also a high incidence of female reproductive tract malformations that are associated with ectopic pregnancy and premature delivery and an increased incidence of uterine fibroids (Stillman 1982, Baird & Newbold 2005). The DES experience also demonstrated the critical importance of timing and dose of exposure to estrogenic chemicals in determining the extent of abnormalities in human female reproductive tract development, with a higher incidence and higher severity of abnormalities after earlier exposure to higher doses (Jefferies et al. 1984).

To date, only two studies have examined the relationship between phytoestrogen exposure and adult female reproductive tract function; in these studies, the exposure was postnatal. One small retrospective study found that women fed soy formula as infants reported longer menstrual bleeding and more dysmenorrhea than women who were not exposed to soy formula (Strom et al. 2001). The second study, which included almost 20,000 women, linked use of soy formula during infancy with a slightly increased risk of early diagnosis of uterine fibroids (D’Aloisio et al. 2010). Although fibroids have not been reported following neonatal phytoestrogen exposure in animal models, these findings are consistent with the observations that rats and humans exposed to DES during developmental periods have an increased incidence of uterine fibroids as adults (Baird & Newbold 2005, Cook et al. 2007).

**Male reproductive tract**

Male fetuses also have adverse effects of developmental exposure to estrogenic chemicals that affect their ability to reproduce. A collaborative cohort study of about 1200 sons of DES-treated women demonstrated an increased incidence of cryptorchidism, epididymal cysts, and inflammation or infection of the testis, particularly when exposure occurred before gestational week 11 (Palmer et al. 2009). These findings are consistent with the reported effects of estradiol in inhibiting insulin-like 3-mediated signaling that is required for testicular descent and estradiol-mediated inhibition of anti-Müllerian hormone activity that is important for Müllerian duct regression (Visser et al. 1998, Cederroth et al. 2007). There are conflicting results in the literature regarding whether DES-exposed males have an increased risk of hypospadias, with several studies indicating an increased risk but other studies not replicating this finding (Brouwers et al. 2010). There is a single report of an approximately fivefold increased risk of hypospadias in sons of mothers who ate a vegetarian diet during pregnancy (North & Golding 2000). Whether this finding is due to phytoestrogen exposure or whether it is caused by concomitant ingestion of other chemicals, e.g. pesticides, that could be more prevalent in a vegetarian diet, is unknown. However, a mouse study regarding the effects of genistein alone or combined with the pesticide vinclozolin suggests that both chemicals can increase the incidence of hypospadias (Vilela et al. 2007).

**Summary**

There is overwhelming evidence in animal models that phytoestrogen exposure can have significant consequences for reproductive health. The effects observed depend on the dose and route of exposure because these parameters impact the final serum level of the bioactive compound. In addition, the timing of exposure is critical in determining phenotypic effects because different tissues have species-specific windows of sensitivity to morphological and functional disruption. These sensitive windows generally begin in the early prenatal period and extend in some cases through adulthood. As more phytoestrogens are recognized or developed as therapeutic compounds, it will be important to examine carefully the effects of these chemicals on reproductive outcomes using animal models that replicate human exposure levels.

Although the effects of phytoestrogens have been evaluated in only a small number of human studies, several of these studies are consistent with findings in animal models after taking into consideration their different developmental timing of specific tissues. Other human studies are more difficult to interpret because of inherent weaknesses in study designs involving large populations of human subjects, difficulties in quantifying phytoestrogen exposures, and an inability to fully control unknown factors in the diet or from the environment that could influence outcomes. There is a need for additional well-designed prospective human studies that limit these variables as much as
possible and focus on environmentally relevant levels of phytoestrogen exposure. Despite their limitations, information gathered from the published human studies combined with the large number of animal studies already available clearly demonstrates that phytoestrogens have the ability to permanently reprogram adult tissue responses after a developmental exposure and that these altered tissue responses are important for reproductive health. These findings should be taken into account when recommendations are made regarding dietary or therapeutic phytoestrogen intake, while keeping in mind the developmentally sensitive time points.

Declaration of interest

The authors declare that there is no conflict of interest that could be perceived as prejudicing the impartiality of the research reported.

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